# **Comparing Explanatory Variables in the Analysis** of Species Composition of a Tallgrass Prairie

#### Joshua M. Brokaw

Department of Botany, Oklahoma State University, Stillwater, OK 74078

Although the relationship between soil characteristics and plant species composition has been well studied, exploratory analyses have been limited by the cost and/or difficulty of analyses. I used Canonical Correspondence Analysis to determine whether a set of potentially important but difficult to measure (or new) soil variables (total C and N, inorganic N, potential net N-mineralization and net nitrification rates, P fractions, and soil textural classes), explain species composition beyond that explained by a more routinely collected (or old) set. Through forward selection, I chose five environmental variables (total C (C), pH, Fe, residual P (P), and Zn) that explained a significant ( $\alpha$ = .05) portion of variation in species composition. However, during a second forward selection using only the old variables, soil organic matter (SOM) and slope were chosen in place of C, and P,. The explanatory power of C, and P, was not significantly greater than that of SOM and slope. In addition, the large number of significant correlations between C, and P, and essential resources suggest that these variables are only indirectly linked to species composition. This study demonstrates that difficult-to-measure variables may be superfluous and that exploratory analyses are likely to choose proxy variables. © 2004 Oklahoma Academy of Science

#### **INTRODUCTION**

The relationship between soil characteristics and species composition is useful for understanding restoration because the success of maintaining or restoring a specific community depends on how management impacts such characteristics (Critchley et al 2002). Canonical Correspondence Analysis (CCA) is a form of direct gradient analysis (Palmer 1993) that is widely used in plant ecology to model the relationship between plant species composition and measured environmental variables (De'ath 2002, Kenkel et al 2002, Lepš and Smilauer 2003, Graffelman and Tuft 2004). Thus, it is a potentially useful technique for restoration ecology. However, exploratory analyses of species composition are limited because soil analysis is often expensive or labor intensive. Fortunately, canonical analyses such as CCA can provide statistics that estimate the proportion of variance of the response data that is accounted for by the explanatory variables (Makarenkov and Legendre 2002), and CCA with variance partitioning allows comparisons between variables or groups of variables (Økland 1994). These properties of CCA allow the comparisons of explanatory variables, and may lead researchers to exclude uninformative or redundant explanatory variables in future analyses.

Nitrogen and P are of particular interest as potential explanatory variables because they are the two most commonly limiting soil nutrients in grasslands (Seastedt et al 1991, Boeye et al 1997, Turner 2003). Nutrient limitation is one of the most important factors affecting plant communities (Grime et al 1997). This might be due to competition for the one most limiting nutrient such as N or differential limitation between species by different nutrients (Koersleman and Meuleman 1996). For example, higher P availability may favor legumes, whereas higher N availability could favor grasses (Janssens et al 1998). The quantity of available nitrogen can have a major influence on species composition and diversity (Mountford et al 1993, Willems et al 1993), and P can control vegetation type and soil organic matter (Smeck 1973). Ratios of these nutrients with organic C may also have substantial effects on vegetation (Koerselman and Meuleman 1996). For instance, a C:N ratio of 14 has been observed to cause a N-mineralization rate that maximized plant productivity, whereas lower ratios caused nitrate leaching and higher ratios caused litter accumulation (Alvarez et al 1998).

Not all variables are equally useful in explaining plant species composition (Palmer et al 2002). Previous exploratory analyses of tallgrass prairie have included N and P gradients that correlate well with productivity in agronomic systems. However, natural communities occur across a wide range of nutrient regimes, and measurements of highly available nutrient pools might not represent the major sources of plant available N and P in unfertilized systems with coevolved niche differentiation (Critchley et al 2002, Schmidt et al 1996). Such systems may rely heavily on the mineralization of organic N and P through microbial activity. Distinctions between mineralization processes, such as ammonification and nitrification, could provide greater explanatory value because plant competition is affected by the form of available N (Schimel et al 1989, Bloom et al 2003). There is substantial temporal heterogeneity and microbial redistribution of P (Hedley et al 1982, Magid and Nielson 1992). Although P-mineralization is difficult to measure, the total P in soils can be divided into inorganic and organic fractions and can be quantified based on levels of bioavailability (Hedley et al 1982). Agriculture-based P measurements quantify the combined total of immediately soluble P plus portions of the more easily extracted, insoluble fractions (Mehlich 1984); however, biologically active P has been found in several fractions (Nichols 1984, Schmidt et al 1996). Fractionation of total P allows for comparisons of pools of P that may be plant available (Abrams and Jarrell 1992) and recalcitrant P forms that may explain community structure through their correlation to soil weathering (Smeck 1973).

It is also important to explore relationships between important environmental gradients. Many variables that display high explanatory power in direct gradient analyses may in fact be serving as proxies for one or many variables that have a more direct, causal relationship with plant species composition. The species composition of the plant community is not only controlled by the initial physical environment but also by the modifications to the physical environment imposed by community succession (Odum 1969). As a result, the causal relationships resulting in simple correlations between species composition and environmental variables may be extremely convoluted. For example, P availability may be controlled by the chemical characteristics of the soil parent material or the chemical characteristics of plant litter (Walker and Adams 1958, Walker et al 1959, Nichols 1984). Phosphorus availability may affect species composition directly by favoring legumes (Walker and Adams 1958, Janssens et al 1998), or it could exert indirect control by affecting N-fixation, N-mineralization, and nitrification (Hue and Adams 1984, Janssens et al 1998). There is strong evidence for control of N-mineralization rates by the C:N ratio (Aulakh et al 2000). Textural classes are strongly correlated with P fractions (O'Halloran et al 1987) and soil organic matter (Hook and Burke 2000). Such correlations by proxy may provide useful generalizations, but it is important to demonstrate causality to apply information provided through direct gradient analyses to management.

My objectives for this study were 1) to estimate the potential of several pools of soil C, N, P, and soil texture to improve explanations of plant species composition in tallgrass prairie and 2) to examine proxy relationships between variables with high explanatory value and other potentially important environmental variables in this community.

### METHODS

All vascular plant species were recorded in 20 permanent 10 m x 10 m plots in the Nature Conservancy's Tallgrass Prairie Preserve during June of 2002 (Palmer et al 2003). The plots were a random sample of grassland plots from a total of 151 that are located at the intersections of the 1 km x 1 km Universal Transverse Mercator Grid. This grid, adopted by the National Imagery and Mapping Agency, consists of two sets of straight, parallel lines, uniformly spaced, each set perpendicular to the other with coordinates measured north and east in meters. These 20 plots have been resurveyed annually beginning in 1998. Species abundance was quantified by estimating percent cover (Palmer et al 2002). Each plot is one sample in the species data for use in CCA (Ter Braak 1986).

For direct gradient analysis, I used estimates of percent slope, aspect, height of grasses, forbs, and woody plants and percent cover of rock, bare ground, and woody plants. I also collected soil samples from each plot. The samples consisted of four combined cores from the top 10 cm of the soil profile (Palmer 1990). I divided each sample into two portions. At Oklahoma



Figure 1. Sequential extraction procedure and functional significance of extracted soil fractions (Pi: inorganic, Po: organic; adapted from Tiessen and Moir 1993)

State University, I measured P fractions (Fig. 1; Tiessen and Moir 1993), inorganic N (Maynard and Kalra 1993), potential net Nmineralization and net nitrification rates (Vinton and Burke 1995), total carbon and nitrogen with a LECO CN 2000 combustion analyzer (Leco, St. Joseph, MI), and soil texture (Gavlak et al 2003). The other portions of these samples were sent to Brookside Labs in New Knoxville, Ohio, to be analyzed for cation exchange capacity, pH, percent soil organic matter (SOM), estimated N-release, soluble S, exchangeable Ca, Mg, K, and Na, and Mehlich III extractable P, Mn, Zn, B, Cu, Fe, and Al (Mehlich 1984). Unlike potential net N-mineralization, estimated N-release is calculated as a function of SOM. I log transformed all variables, excluding pH, that were derived from these soil analyses (Palmer 1990). In addition, I included easting, northing, and sampling date for each plot sampled in the analysis. For the purposes of discussion I will refer to total C and N, inorganic N, mineralization and nitrification rates, P fractions, and soils textural classes as the new environmental variables. All other environmental data constitute the old variables.

Direct gradient analysis (Palmer 1993) and variance partitioning (Borcard et al 1992, Økland 1994, 2003) with the environmental variables and species data, were performed using CCA through CANOCO for Windows software (Ter Braak and Smilauer 1998). I chose to square-root transform the species data in order to reduce the effects of dominant species and down weight rare species prior to analysis. Because the number of environmental variables collected was greater than the number of samples, stepwise forward selection was used to choose the environmental variables from the full set that explained the greatest amount of variation in plant species composition within the samples (Ter Braak 1988b, Hallgren et al 1999). Stepwise forward selection is a semi-automated procedure performed through CANOCO. Environmental variables are added, one at a time, until

35

no other variables significantly explain residual variation in species composition. However, the user is allowed to make choices about inclusion of variables. In addition, I used stepwise forward selection to choose environmental variables from only the old set of variables to detect those variables that are potentially interchangeable with new variables. Variables with *P* values less than 0.05 derived through Monte-Carlo permutations tests with 999 permutations were chosen for further analyses.

I used variance partitioning (Økland 2003) to evaluate the redundancy in explanatory value of a set of old (O) and two new (N<sub>1</sub> and N<sub>2</sub>) variables chosen through forward selection. The CCA of all selected variables measures the total inertia (TI) of the variation in plant species composition and the inertia explained by the union of the 3 sets ( $N_1 \cup N_2 \cup O$ ; n = 3). Inertia is analogous to the variance in a data set (Palmer 2004). I divided the TI explained by  $N_1 \cup N_2 \cup O$  into  $2^n - 1 = 2^3 - 1 = 7$  components in order to determine the relative importance of the three sets of variables, separately and jointly. I quantified the inertia uniquely explained by each set  $(N_1 | N_2 \cup O)$ ,  $N_{a} | N_{1} \cup O$ , and  $O | N_{1} \cup N_{a}$ ) with three partial CCAs in which I entered one set as environmental variables and the two remaining sets as covariables. A partial CCA is a CCA in which the effects of certain environmental variables (covariables) are removed. For example N,  $|N_0 \cup O|$  represents the variation explained by N<sub>1</sub> not included in N<sub>2</sub> $\cup$ O. The redundant portions of two or more data sets are called the intersections. I quantified the intersections between all three sets  $(N_1 \cap N_2 \cap O)$  and each pair of sets  $(N_1 \cap N_2 \mid O)$ ,  $N_1 \cap O | N_2$ , and  $N_2 \cap O | N_1$ ) indirectly. For example,  $N_1 \cap N_2 | O$  is the difference between  $N_1 \cup N_2 \mid O$  and the sum of  $N_1 \mid N_2 \cup O$ and  $N_{2} \mid N_{1} \cup O$ , and  $N_{1} \cap N_{2} \cap O$  is the difference between  $N_1 \cup N_2 \cup O$  and the sum of  $N_1 | N_2 \cup O, N_2 | N_1 \cup O, O | N_1 \cup N_2, N_1 \cup N_2 | O,$  $N_{2} \cup N_{1} \mid O$ , and  $O \cup N_{1} \mid N_{2}$ . I used variance partitioning to test the null hypothesis: the new environmental variables do not explain

variation in plant species beyond that which is explained by the old environmental variables (Ter Braak 1986, 1987, 1988a).

In addition to CCA, I compared all environmental variables by using Pearson correlations with SPSS (2001). As I displayed these correlation coefficients to assess the strength, rather than the significance of these relationships, I did not correct for the multiple correlation (Legendre and Legendre 1998).

#### RESULTS

Forward selection chose environmental variables in the order total C (C) (P = 0.001), Mehlich III extractable Fe (Fe) (P = 0.001), pH (P = 0.002), residual P ( $P_{..}$ ) (P = 0.031), and Mehlich III extractable Zn (Zn) (P =0.041). The remaining variation in plant species composition could not be significantly explained ( $\alpha = 0.05$ ) with the available set of environmental variables. The selected variables explained 41% of the total inertia of the species data.  $C_{t}$  and  $P_{r}$  were chosen from the new set of variables. Forward selection from only the old variables set resulted in the following selection: SOM (P = 0.001), Fe (P = 0.001), pH (P = 0.001), slope (P = 0.044), and Zn (P = 0.044). These variables also explained 41% of the total inertia of the species data.

Variance partitioning shows that 37.9% of the variation explained by  $C_t$ , Fe, pH,  $P_r$ , and Zn was uniquely explained by the new variables  $C_t$  and  $P_r$  (Fig. 2).  $C_t$  uniquely accounts for 20.1% of the explained variation, and  $P_r$  accounts for 15.0% of the explained variation. I do not display variance partitioning between the selected new variables and the alternate old variables (SOM and slope) because the partial CCAs necessary to produce N | O and O | N had *P* values of 0.16 and 0.21, respectively ( $\alpha = 0.05$ ).

When testing with two-tailed tests of significance,  $C_t$ , Fe, pH,  $P_r$ , and Zn were significantly correlated ( $\alpha = 0.05$ ) with 25, 12, 18, 17, and 4, respectively, of the unused environmental variables in the data set.



Figure 2. Partitioning of variation in plant species composition explained by the soil variables total C, Mehlich III extractable Fe, pH, residual P, and Mehlich III extractable Zn. Variation explained by Fe, pH, and Zn is grouped as one unit.  $\cap$  = intersection.

Total N was the soil nutrient variable most highly correlated with both C<sub>1</sub> and P<sub>1</sub> (Table 1). All variables were significantly correlated with C<sub>t</sub> except NaHCO<sub>3</sub> P<sub>o</sub>, NaOH P<sub>o</sub>, and Na. Most soil nutrients were significantly correlated with P<sub>r</sub> except the labile forms of N and P and Na. C, was significantly correlated with all variables associated with biomass and soil type. P, was significantly correlated with all texture variables, SOM, and forb height. C, was significantly correlated with all inorganic fractions of soil P and organic soil P extracted with concentrated HCl. P<sub>r</sub> was significantly correlated with the other recalcitrant forms of soil P (HCl extractable P and NaOH extractable organic P). The correlation coefficient between  $P_r$  and slope was 0.438 (P = 0.053).

## DISCUSSION

Based on the results of variance partitioning, the new environmental variables were unable to explain variance in species composition beyond that explained by the old Table 1. Pearson correlation coefficientsbetween total C, residual P, and soil nutri-ents, soil texture, and site characteristics.

|                      | Total C      | Residual P |  |
|----------------------|--------------|------------|--|
| Total N              | _0.977**     | _0.679**   |  |
| NO3                  | _0.464*      | -0.017     |  |
| NH                   | _0.534*      | _0.179     |  |
| Resin P              | _0.664**     | _0.353     |  |
| NaHCO <sub>3</sub> P | _0.455*      | _0.240     |  |
| NaOH P               | _0.512*      | _0.392     |  |
| 1M HCl P             | _0.542*      | _0.485*    |  |
| cHCl P <sub>i</sub>  | _0.866**     | _0.781**   |  |
| NaHCO <sub>3</sub> P | -0.361       | -0.004     |  |
| NaOH P               | _0.398       | _0.546*    |  |
| cHCl P               | _0.899**     | _0.790**   |  |
| K                    | _0.583**     | _0.551**   |  |
| SO <sub>4</sub>      | _0.808**     | _0.632**   |  |
| Ca                   | _0.868**     | _0.633**   |  |
| Mg                   | _0.827**     | _0.650*    |  |
| Na                   | _0.234       | _0.186     |  |
| Sand                 | -0.835**     | -0.778**   |  |
| Silt                 | _0.758**     | _0.689**   |  |
| Clay                 | _0.892**     | _0.693**   |  |
| Soil organic         | _0.959**     | _0.713**   |  |
| Bare ground          | $-0.496^{*}$ | -0.337     |  |
| Grass height         | _0.588**     | _0.259     |  |
| Forb height          | _0.523*      | _0.473*    |  |
|                      |              |            |  |

\*\* Significant at the 0.01 level (2-tailed)

\* Significant at the 0.05 level (2-tailed)

environmental variables. The low level of redundancy found between  $C_t$ , Fe, pH,  $P_r$ , and Zn suggests that  $C_t$  and  $P_r$  could provide additional, significant explanation of species composition. However, partial CCAs comparing  $C_t$  and  $P_r$  to SOM and slope show that differences between the new variables and alternate old variables are not significant.

The lack of significance is primarily due to the strong correlation between  $C_t$  and SOM.  $C_t$  represents pools of inorganic C, which may be found in abundance in limestone soils, in addition to the organic C rep-

Proc. Okla. Acad. Sci. 84: pp 33-40 (2004)

resented by SOM. However, organic C also tends to be higher in limestone soils. By combining two of the common characteristics of soils forming from limestone parent material, high inorganic C and high organic content,  $C_t$  could provide a marginal increase in explanatory power over variables representing organic content and Ca content in plant communities that are highly influenced by soil type such as the boundary between cross-timbers and prairie (Francaviglia 2000). Unfortunately, the small sample size in this study is inadequate for outlining the differences in such highly correlated variables.

P<sub>r</sub> was not so highly correlated with any single variable in the available set, and its correlation with slope was only marginally significant. However, interpreting its value in explaining plant species composition is problematic. P<sub>r</sub> tends to be correlated, although weakly, with many of the same environmental variables as C<sub>t</sub>. This is likely because recalcitrant P is also associated with a limestone parent material (Schlesinger 1997). The explanatory value of P<sub>v</sub> beyond that of the other selected variables, though statistically significant, shows no discernable relationship with the ecological traits of the species variables. In addition, P, almost certainly functions as a proxy variable because the pool of soil P represented by P<sub>r</sub> is not bioavailable without extensive weathering. As a proxy variable P<sub>r</sub> is likely related to soil type and extent of weathering (Smeck 1973). A stronger causal connection between P<sub>a</sub> and species responses is still needed to interpret the role of P<sub>v</sub> in controlling species composition. It is also possible that the significant explanatory value of P<sub>r</sub> is an artifact of a low sample size. Unfortunately, the high cost of P fractionation suggests that P<sub>r</sub> is not likely to play a substantial role in the future of direct gradient analysis.

C<sub>t</sub> and P<sub>r</sub> are strong examples of the use of proxy variables in direct gradient analysis. Unlike indirect gradient analysis, direct gradient analyses, such as CCA, constrain the scores of the response variables to be linear combinations of explanatory variables. The variation in species composition of the samples is represented in terms of the chosen explanatory variables, but a causal relationship cannot be guaranteed. Environmental gradients, such as C<sub>1</sub>, SOM, Ca, and pH, tend to be strongly correlated with patterns in plant species composition, but it is difficult to unravel the causal relationships responsible for this correlation. It is likely that most if not all of these variables serve as proxies for environmental conditions contributing to the spatial arrangement of species within the prairie community. Resources traditionally recognized to limit plant growth include space, light, water, and nutrients. C<sub>t</sub> and P<sub>r</sub> are correlated with environmental variables associated with the availability of space, light, and water, and C, is correlated with plant available nutrients (Table 1). The correlation of C, and P, with such a large number of potentially influential environmental variables suggests that the relationships between C, and P, and plant species composition are probably not

It is not surprising that proxy variables representing multiple factors affecting species composition were chosen through stepwise forward selection because I intentionally chose variables that explain the greatest amount of variation in the species data. Future research intended to identify more direct relationships could focus on more homogeneous systems, such as only on communities within a single soil type. In contrast, when studying a more heterogeneous system, such as the prairiecrosstimbers continuum, proxy variables should become more dominant by explaining differences between forest and grassland vegetation. Due to the interdependence of organisms and multiple biogeochemical cycles, proxy variables will likely continue to play an important role in exploratory analysis.

directly causal.

## ACKNOWLEDGMENTS

I thank M. Palmer, R. Tyrl, and A. Cross for their suggestions on this manuscript; B. Hamilton and The Nature Conservancy for access to sites at The Tallgrass Prairie Preserve; and M. Palmer, P. Earls, K. Ray, and A. Brokaw for assisting me with field and laboratory work. I especially thank M. Palmer for supervising the collection of plant species data and providing valuable assistance during data analyses. This work represents a portion of a thesis submitted in partial fulfillment of the requirements for a Master of Science degree and was funded by the James K. McPherson Memorial Fund.

### REFERENCES

- Abrams MM, Jarrell WM. 1992. Bioavailability index for phosphorus using ion exchange resin impregnated membranes. Soil Sci Soc Am J 56:1532-1537.
- Alvarez G, Chaussod R, Loiseau P, Delpy R. 1998. Soil indicators of C and N transformations under pure and mixed grass-clover swards. Eur J Agron 9:157-172.
- Aulakh MS, Khera TS, Doran JW. 2000. Mineralization and denitrification in upland, nearly saturated and flooded subtropical soil. Biol Fert Soils 31:168-174.
- Bloom AJ, Meyerhoff PA, Taylor AR, Rost TL. 2003. Root development and absorption of ammonium and nitrate from the rhizosphere. J Plant Growth Regul 21:416-431.
- Boeye D, Verhagen B, Van Haesebroeck V, Verheyen RF. 1997. Nutrient limitation in species-rich lowland fens. J Veg Sci 8:415-424.
- Borcard D, Legendre P, Drapeau P. 1992. Partialling out the spatial components of ecological variation. Ecology 73:1045-1055.
- Critchley CNR, Chambers BJ, Fowbert JA, Sanderson RA, Bhogal A, Rose SC. 2002. Association between lowland grassland plant communities and soil properties. Biol Conserv 105:199-215.
- De'ath G. 2002. Multivariate regression trees: a new technique for modeling species-environment relationships. Ecology 83:1105-1117.
- Francaviglia RV. 2000. The cast iron forest: a natural and cultural history of the north american cross timbers. Austin (TX): University of Texas Press. 276 p.
- Gavlak RG, Horneck DA, Miller RO, Kotuby-Amacher J. 2003. Plant, soil and water reference methods for the western region. 2nd ed. Western Region Extension Publication [on-line]. WREP-125. p 129-131. Available from http://cropandsoil. oregonstate.edu/nm/WCC103/Soil\_Methods.htm. (Accessed November 9, 2004).
- Graffelman J, Tuft R. 2004. Site scores and conditional biplots in canonical correspondence analysis. Environmetrics 15:67-80.

- Grime JP, Thompson K, Hunt R, Hodgson JG, Cornelissen JHC, Rorison IH, Hendry GAF, Ashenden TW, Askew AP, Band SR, Booth RE, Bossard CC, Cambell BD, Cooper JEL, Davison AW, Gupta PL, Hall W, Hand DW, Hannah MA, Hillier SH, Hodkinson DJ, Jalili A, Liu Z, Mackey JML, Matthews N, Mowforth MA, Neal AM, Reader RJ, Reiling K, Ross-Fraser W, Spencer RE, Sutton F, Tasker DE, Thorpe PC, Whitehouse J. 1997. Integrated screening validates primary axes of specialization in plants. Oikos 79:259-281.
- Hallgren E, Palmer MW, Milberg P. 1999. Data diving with cross validation: an investigation of broadscale gradients in Swedish weed communities. J Ecol 87:1037-1051.
- Hedley MJ, Stewart JWB, Chauhan BS. 1982. Changes in inorganic and organic soil phosphorus fractions induced by cultivation practices and laboratory incubations. Soil Sci Soc Am J 46:970-976.
- Hook PB, Burke IC. 2000. Biogeochemistry in a shortgrass landscape: control by topography, soil texture, and microclimate. Ecology 81:2686-2703.
- Hue NV, Adams F. 1984. Effect of phosphorus level on nitrification rates in three low-phosphorus ultisols. Soil Sci 137:324-331.
- Janssens F, Peeters A, Tallowin JRB, Bakker JP, Bekker RM, Fillat F, Oomes MJM. 1998. Relationship between soil chemical factors and grassland diversity. Plant Soil 202:69-78.
- Kenkel NC, Derksen DA, Thomas AG, Watson PR. 2002. Multivariate analysis in weed science research. Weed Science 50:281-292.
- Koerselman W, Meuleman AFM. 1996. The vegetation N:P ratio: a new tool to detect the nature of nutrient limitation. J Appl Ecol 33:1441-1450.
- Legendre P, Legendre L. 1998. Numerical ecology. 2nd English ed. Amsterdam (NA): Elsevier. 853 p.
- Lepš J, Šmilauer P. 2003. Multivariate analysis of ecological data using CANOCO. Cambridge (UK): Cambridge University Press. 282 p.
- Magid J, Nielsen NE. 1992. Seasonal variation in organic and inorganic phosphorus fractions of temperate climate sandy soils. Plant Soil 144:155-165.
- Makarenkov V, Legendre P. 2002. Nonlinear redundancy analysis and canonical correspondence analysis based on polynomial regression. Ecology 83:1146-1161.
- Maynard DG, Kalra YP. 1993. Nitrate and exchangeable ammonium nitrogen. In: Carter MR, editor. Soil sampling and methods of analysis. Boca Raton (FL): Lewis Publishers. p 25-38.
- Mehlich A. 1984. Mehlich 3 soil test extractant: a modification of Mehlich 2 extractant. Commun Soil Sci Plant 15:1409-1416.
- Mountford JO, Lakhani KH, and Kirkham FW. 1993. Experimental assessment of the effects of nitrogen addition under hay-cutting and aftermath grazing on the vegetation of meadows on a Somerset peat moor. J Appl Ecol 30:321-332.
- Nichols JD. 1984. Relation of organic carbon to soil properties in the southern Great Plains. Soil Sci Soc Am J 48:1382-1384.
- Odum EP. 1969. The strategy of ecosystem development. Science 164:262-270.

Proc. Okla. Acad. Sci. 84: pp 33-40 (2004)

- O'Halloran IP, Stewart JWB, Kachanoski RG. 1987. Influence of texture and management practices on the forms and distribution of soil phosphorus. Can J Soil Sci 67:147-163.
- Økland RH. 1994. Canonical correspondence analysis with variation partitioning: some comments and an application. J Veg Sci 5:117-126.
- Økland RH. 2003. Partitioning the variation in a plotby-species data matrix that is related to *n* sets of explanatory variables. J Veg Sci 14:693-700.
- Palmer MW. 1990. Spatial scale and patterns of species-environment relationships in hardwood forest of the North Carolina piedmont. Coenoses 5:79-87.
- Palmer MW. 1993. Putting things in even better order: the advantages of Canonical Correspondence Analysis. Ecology 74:2215-2230.
- Palmer MW. 2004. Ordination methods for ecologists [online]. Available from: http://www.okstate.edu/ artsci/botany/ordinate/. (Accessed November 9, 2004).
- Palmer MW, Arévalo JR, Cobo MC, Earls PG. 2003. Species richness and soil reaction in a northeastern Oklahoma landscape. Folia Geobot 38:381-389.
- Palmer MW, Earls PG, Hoagland BW, White PS, Wohlgemuth T. 2002. Quantitative tools for perfecting species lists. Environmetrics 13:121-137.
- Schlesinger WH. 1997. Biogeochemistry: an analysis of global change. 2nd ed. New York (NY): Academic Press. 588 p.
- Schimel JP, Jackson LE, Firestone MK. 1989. Spatial and temporal effects on plant-microbial competition for inorganic nitrogen in a California annual grassland. Soil Biol Biochem 21:1059-1066.
- Schmidt JP, Buol SW, Kamprath EJ. 1996. Soil Phosphorus dynamics during seventeen years of continuous cultivation: fractionation analyses. Soil Sci Soc Am J 60:1168-1172.
- Seastedt TR, Briggs JM, Gibson DJ. 1991. Controls of nitrogen limitation in tallgrass prairie. Oecologia 87:72-79.
- Smeck NE. 1973. Phosphorus: an indicator of pedogenic weathering processes. Soil Sci 115:199-206.
- SPSS. [computer program]. 2001. SPSS for Windows. Release 11.0.1 Standard version. Chicago (IL): SPSS Inc.

- Ter Braak CJF. 1986. Canonical Correspondence Analysis: a new eigenvector technique for multivariate direct gradient analysis. Ecology 67:1167-1179.
- Ter Braak CJF. 1987. The analysis of vegetation-environment relationships by canonical correspondence analysis. Vegetatio 69:69-77.
- Ter Braak CJF. 1988a. CANOCO—an extension of DECORANA to analyze species-environment relationships. Vegetatio 75:159-160.
- Ter Braak CJF. 1988b. Partial canonical correspondence analysis. In: Bock HH, editor. Classification and related methods of data analysis. Amsterdam (NA): North-Holland. p 551-558.
- Ter Braak CJF, Šmilauer P. 1998. CANOCO reference manual and user's guide to CANOCO for Windows: software for Canonical Community Ordination. version 4. Ithaca (NY): Microcomputer Power. 352 p.
- Tiessen H, Moir JO. 1993. Characterization of available P by sequential fractionation. In: Carter MR editor. Soil sampling and methods of analysis. Boca Raton (FL): Lewis Publishers. p 75-86.
- Turner BL, Chudek JA, Whitton BA, Baxter R. 2003. Phosphorus composition of upland soils polluted by long-term atmospheric nitrogen deposition. Biogeochemistry 65:259-274.
- Vinton MA, Burke IC. 1995. Interactions between individual plant species and soil nutrient status in shortgrass steppe. Ecology 76:1116-1133.
- Walker TW, Adams AFR. 1958. Studies on soil organic matter: I. Influence of phosphorus content of parent materials on accumulations of carbon, nitrogen, sulfur, and organic phosphorus in grassland soils. Soil Sci 85:307-213.
- Walker TW, Thapa BK, Adams AFR. 1959. Studies on soil organic matter: 3. Accumulation of carbon, nitrogen sulfur, organic and total phosphorus in improved grassland soils. Soil Sci 87:135-141.
- Willems JH, Peet RK, Bik L. 1993. Changes in chalk grassland structure and species richness resulting from selective nutrient additions. J Veg Sci 4:203-212.

Received: May 4, 2004; Accepted November 16, 2004